

Assessment of Environmental Geochemistry of Lead-Zinc Mining at Ishiagu Area, Lower Benue Trough, Southeastern Nigeria

Odika, P.O.¹, Anike, O.L.¹, Onwuemesi, A.G.¹, Odika, N.F.¹ & Ejeckam, R.B.²

¹ Department of Geological Sciences, Nnamdi Azikiwe University, Awka

² Environment Canada, Environmental Protection Operations, Winnipeg, MB, Canada

Correspondence: Odika, P.O., Department of Geological Sciences, Nnamdi Azikiwe University, PMB 5025 Awka, Nigeria

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Abstract

Mining activities have long been recognized as a major source of environmental contamination associated with heavy metals and metalloids. This study evaluated the relationship between the occurrence and mining of lead-zinc sulphide ores at Ishiagu, Nigeria, and heavy metal and metalloid contamination. A comparative study of two zones in the area, with and without mining activities was also made. Water, soil, stream sediment and ore samples were analyzed, after acid digestion, using atomic absorption spectrophotometer (AAS). The concentration levels of seven heavy metals and a metalloid namely Pb, Cu, Ni, Zn, Mn, Co, Cd and As were evaluated. While the highest concentration levels of As, Co and Pb (5.20 mg/l, 0.54 mg/l and 3.40 mg/l respectively) were found in water, those of Ni and Mn (2.26 mg/l and 5.48 mg/l respectively) occurred in soil. For Cu and Zn, highest levels of concentration (2.80 mg/l and 0.41 mg/l respectively) occurred in stream sediments. The variations in the concentration levels of these elements in varying geologic media (soil, water and sediment) indicate influence of rock types, human activities and media physicochemical characteristics. Geostatistical analyses using QQPLOT, semivariogram and kriging showed normal distribution of these elements. Distribution and dispersion patterns of the heavy metals indicated increase in concentration levels in the local stream flow direction. Pb, Cu, As, Cd, Mn, and Ni concentrations had reached pollutant levels in water based on WHO standards, while Zn level is below. Since the local people use untreated surface water and groundwater for drinking and other domestic purposes, soil for farming and lead for cosmetics, long term exposure poses significant health risk for humans, animals and plants.

Keywords: Lead-Zinc mineralization, igneous rock quarrying, traces elements, environmental, pollution/contamination

1. Introduction

There is usually an interaction in the rock-water-air-life system, which gives rise to a wide range of chemical characteristics in the surface environment (Thornton, 1983). Most of these elements are essential such as Mn, Fe, Zn, Mg and Ca, while others are toxic (including As, Pb and Hg) to humans, animals, and plants (Chellan & Sadler, 2015; Hajeb et al., 2014; Thornton, 1983; Carla, 2002; Carter & Stewart, 2000; Chorover et al., 2007). The toxicity depends mostly on dosage, availability and accessibility. The toxic metals and metalloids have often been described as heavy, with specific gravity greater than 5.0. Examples of these metals and metalloids are Pb, Cd, As, Co, Cu, Zn, Ni and Mn (Edelstein & Ben-Hur, 2018; Cooksey, 2012; Babula et al., 2008; Tamas et al., 2014). They cause environmental pollution from a number of sources. The accumulation of heavy metals in agricultural soils and water resources is of increasing concern due to food safety issues, potential health risks and contamination impacts on the ecosystems.

Lead-zinc (Pb-Zn) sulphide mining activities in the Ishiagu area in the lower Benue Trough, southeastern Nigeria have a potential significant impact on the quality of soil and water resources and, by extension, animal health and agricultural produce within the environment. Mining activities date back to the 1960s (NGSA, 2010). Previous investigations have centered largely on flotation studies (Onyemaobi, 1990), hydrometallurgy (Adebayo et al., 2003), mineralogical characterization (Olubambi et al., 2008), physicochemical properties of mine pit water (Aroh et al., 2007), geoelectric exploration (Orajaka et al., 1988) and geologic setting (Ezepue, 1984). Also, Ezepue (1985), Olubambi *et al.* (2008), Eze & Uko (2003), Aroh (2003) and Aroh et al. (2006) determined the mineralogy and geochemistry of epithermal galena, siderite and trace metal status of streams receiving acid mine drainage (AMD). According to Kovacs et al. (2012), neutralized acidic mine drainage and abandoned tailing along river and stream banks pose risk of heavy metal remobilization as a secondary source of pollution.

Previous studies suggest that Pb-Zn ores around the study area occur at shallow depths ranging from 10m to 25m, and mostly as open space fillings of a series of steeply dipping fractures, which cross cut the regional fold axis of the Abakaliki anticlinorium (Orajaka et al., 1988). The ores are epigenetic, massive and strata bound, and are typically formed by hydrothermal processes (Olubambi et al., 2008; Nwabufor-Ene, 1993) within the sedimentary host rocks. The deposits are hosted by carbonaceous shale and contain mostly Fe, Al, Mg, Si, S, Zn, Cu, Ni, Cd, Co, Mn, Pb and As. The mineralization comprises massive sulfides containing mainly sphalerite, galena, siderite, quartz and traces of pyrite and copper bearing minerals (Oha et al., 2017; Ezepue, 1984). Although there have been several geophysical, quantitative and chemical studies of the deposits (Adebayo et al., 2003; Olubambi et al., 2008; Orajaka et al., 1988, Ezepue, 1984, Aroh et al., 2006), the concentration distribution pattern of contaminants and potential impacts of abandoned mines and mine wastes are yet to be adequately assessed and addressed (Aroh, et al., 2006, 2007, Foley, 2002, 2011, Cidu, 2002, 2007, 2009, Aroh, 2003).

Further, proper environmental regulations and enforcement for mining activities are not well promoted. In addition, loss of lives and property resulting from mining activities have not been properly documented or addressed by the relevant authorities. Chemical contamination or pollution of water and soil systems at both abandoned and active mining sites demands rigorous investigation and proper attention. For instance, Mejia (2015) reported that in 2010, Zamfara State in northern Nigeria witnessed lead poisoning from artisanal gold mining, with an average concentration in the local river as high as 1270mg/l. This resulted in the death of more than 300 people, who were mostly artisans, including children, youth and women. Other environmental concerns involve the effect of extensive past magmatic activities and the occurrence of intrusive and extrusive rocks in the Benue Trough (Etuk et al., 2008). These igneous rocks occur over a distance of 500km from Ishiagu in the Lower Benue Trough to Zurak in the Upper Benue Trough with the number of occurrences increasing from Zurak towards the south (Ezepue, 1984, Etuk et al., 2008). Obiorah & Umeji (1995) noted that there was emplacement of numerous mafic intrusive and alkaline/calc-alkaline lava and tuffs associated with the Santonian deformation. The Ishiagu area is characterized by both commercial and artisanal mining and quarrying. The current mining methods are open-pit, involving the excavation of large quantities of host rock in order to extract the desired ore. The ore is processed by pulverization and flotation and the concentrates are usually shipped to smelters locally or exported overseas. The local population also extracts these deposits for domestic purposes such as cosmetics or medical. In addition, artisanal quarrying of igneous rock avail the locals with aggregates for domestic use and local trading. These activities have over a long period served many in the community and visitors as a means of livelihood.

In this work, two areas Ishiagu and Iheututu were studied. This paper presents the results of environmental geological and geochemical studies conducted around mining and quarry sites within these zones to assess the effects of these activities. Elements concentration levels in different media with respect to mining and quarrying sites demand investigation. A combination of geological, geochemical and petrographic data can reveal the source, nature and extent of elements concentrations (Mielka et al., 2000, Facchinelli et al., 2001, Dold, 2005). The methods employed for this study are capable of identifying the sources of contaminants and pollutants. The study will also establish baseline data for soil, stream sediment and surface and groundwater for the area against which future research results could be measured.

2. Study Area

The study area lies between latitudes 5.52 – 6.00 N and longitudes 7.30 – 7.36 E (Figure 1). Locally, the relief is undulating with rolling hills and hillocks that rise up to 93m above sea level. Egboka & Uma (1986) and Hall & Medler (1975) observed that the climatic setting is as varied as savannah conditions in the northern part with stunted trees, derelict woodland, few shrubs and dispersed large trees, and the tropical rain forest in the south with luxuriant evergreen trees. The rain forest vegetation is dominant. There are two major seasons; the wet season (from March to October) and the dry season (from October to February/March). These seasons arise from the two prevailing winds at different periods of the year; the dry harmattan wind from the Sahara desert and the marine wind from the Atlantic Ocean. The temperature ranges from 20° to 38°C in the dry season and from 16° to 28°C in the rainy season.

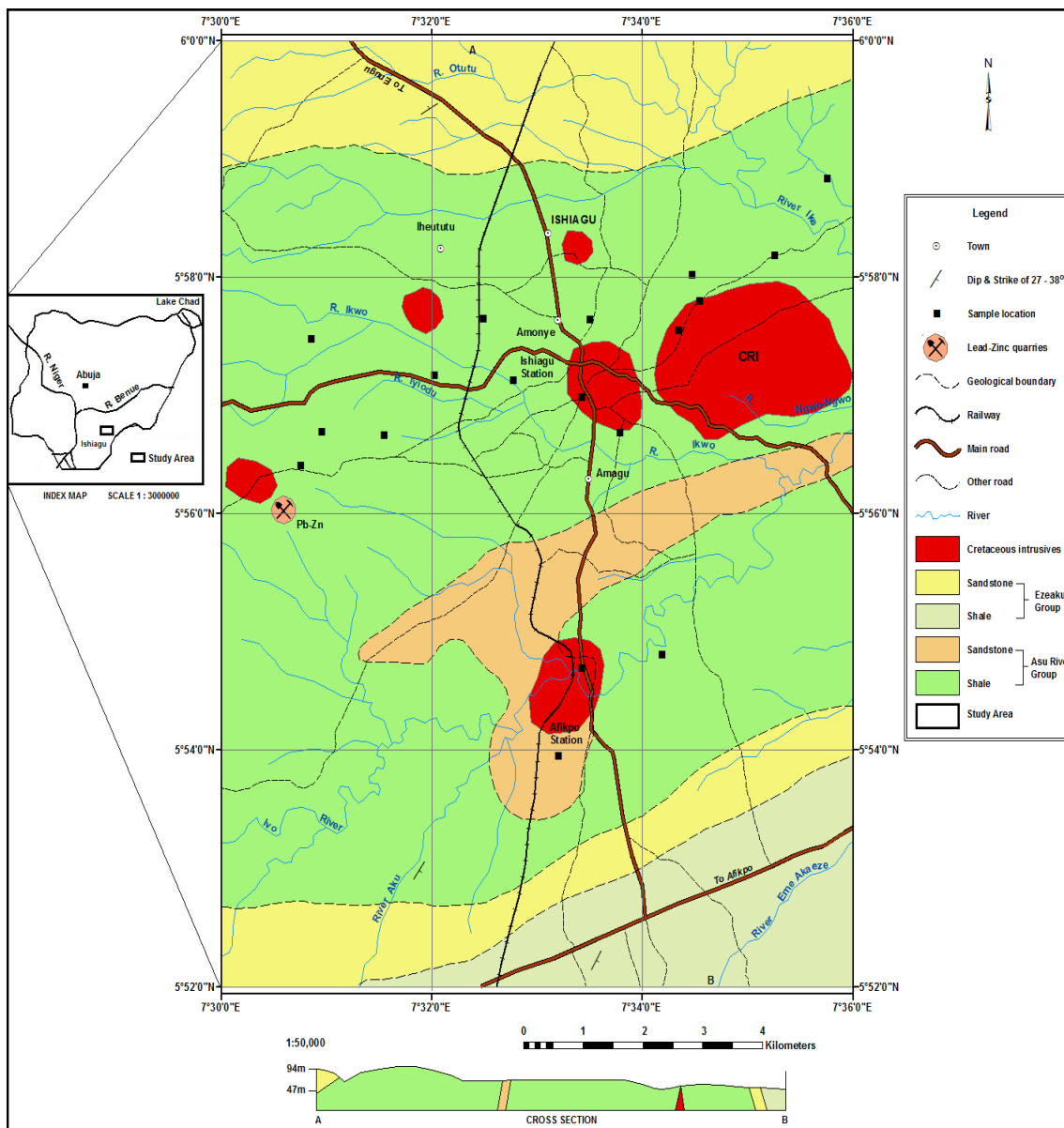


Figure 1. Geological and location map of the study area, inset is map of Nigeria with a black box showing the study area (modified after Oha, 2017)

The area is predominantly underlain by the shale of the Asu River Group of the Albian age (Figure 2). The shale grades unconformably into the younger sequence of sandstone and shale units of the Eze-Aku Formation of Late Cenomanian-Turonian age. Field observations show epigenetic massive sulphide ore deposits hosted by folded and faulted shales. The shales are predominantly black or dark grey, finely laminated and fissile. They are associated with intrusive igneous rocks. The younger sandstone unit of the Eze-Aku Formation underlies the northern and southern parts of the area. This unit is generally greyish white (turned brown by weathering), fine-grained and poorly sorted. It strikes ENE –NE with dips of 27° – 38° NW and SE, and rest unconformably on the Albian Asu River shale. The shale unit of the Eze-Aku Formation is grey, carbonaceous, thinly to thickly bedded, flaggy and fissile on weathering.

The ore deposits in the study area and elsewhere in the Benue Trough have been analysed by previous workers. Northeast of the study area the ores comprise predominantly Pb-Cu-Mn-Ni sulfides, whereas, those in the southwest area show strong enrichment of Pb-Zn-As. Ezeh (2007), reported that in Enyigba and Abakaliki areas (close to the study area), ores contain predominantly Pb-Zn sulfides, while in Uburu and Okposi zones (also close to the study area), the ores are dominated by mostly Cu-rich sulfides. The simple sulfide mineral assemblages of Ishiagu ore deposits consist of galena,

sphalerite, pyrite, marcasite and chalcopyrite. Combined lead plus zinc grades in Ishiagu deposits may constitute up to 15 % of the veins, while siderite may be present up to 80% and quartz 5% (Olubambi et al., 2008, Okafor & Uwadiale, 1997, Ezepue, 1984). Trace amounts of other sulphide and sulfosalt minerals make up the balance. Within a given lead-sulphide lode, mineralogy of the individual deposit may be quite variable. Secondary minerals of the ores are anglesite, smithsonite, cerrusite, malachite, sulfur, gypsum, hydrous iron oxide minerals, hematite, greenockite, Malachite and azurite (Olubambi et al., 2008).

It should be noted that the local population has worked these deposits for domestic uses, while numerous mining companies have extracted some of the ores for commercial purposes, starting from 1962 up to the present. Other long term existing land use activities include rock quarrying (both commercial and artisanal), farming and local herding.

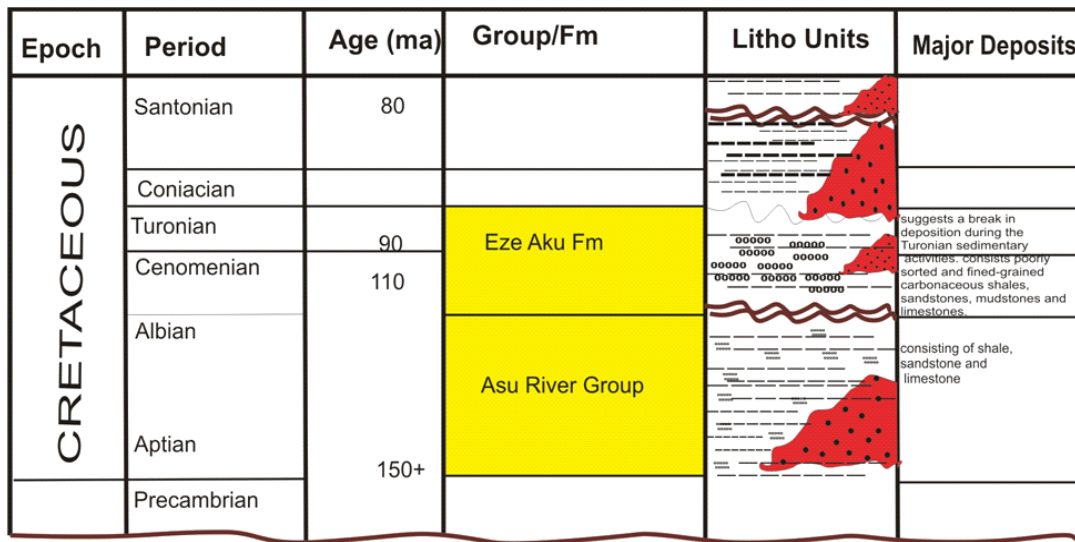


Figure 2. Lithostratigraphic succession of part of the southern Benue Trough covering the study area (boxes in yellow are the formations found in the study area)

3. Material and Methods

Twenty-six out of a total number of thirty-six samples were analysed; eleven samples from water, six samples from soil, three stream sediment samples, three tailing samples and three igneous rock samples. Samples were collected widespread in the study area and as dictated by locations of water sources, mineralizations and intrusives, including from around an ore mining site and an igneous rock quarry. Sample collection, preservation and chemical analyses were done based on modern methods for environmental sample preparation and analysis according to Santelli et al. (2004).

Collection and preparation of samples

Water samples were collected from streams, rivers, a spring, a hand-dug well and boreholes to assess the concentration levels of contaminants in different environmental media. For each sample, about 250ml was collected in hard polyethylene bottles. The bottles were cleaned using metal-free 0.1M HNO₃, and rinsed thoroughly with the sample water at the point of collection. Values of the pH, temperature, TDS, and conductivity were determined and recorded in the field. The samples were refrigerated at 4oC and analyzed within four (4) days of collection in accordance with standard specifications (Santelli et al., 2004). Soil samples were collected from the bottom of holes 5cm to 25cm deep depending on location and induration, using plastic picks and trowels. Stream sediment samples were collected from river and stream beds using an auger. At each location, about 100g of fine fractions were collected and labeled. The damp stream sediment and soil samples were air dried for three days. Sun or oven drying was not used in order to prevent chemical reactions or decomposition of elements or compounds.

Analysis of samples

The dried samples were further disaggregated (with agate mortar) and sieved to 150µm size fraction (using non-metallic Endecotts electronic shaker, M-073038). 1.0 g of well mixed sample, weighed with Mettler Toledo MPL 203, of each was digested using 10 ml of aqua regia, 5ml of HF and 25 ml of boric acid solution. The analytes were further filtered and analyzed for total metal concentrations of Pb, Zn, Cd, Co, Cu, As, Mn, and Ni using the AAS. Sample decomposition was done to bring the analyte concentrations of the solid samples within the working range of the analytical technique selected

(AAS); that is, to aid the reduction of chemical manipulation in scale and complexity, and for minimal quantities of reagents to be used. Thus, 100ml of well mixed sample each was mixed with 20 ml of aqua regia (HNO_3 (150ml)) + HCl (450ml), deionised water + 25ml H_3BO_3 and 5ml HF) to keep the elements in analytic state. Hydrofluoric acid was added to dissolve any suspended and siliceous material (such as Ca, Si) and boric acid to neutralize the HF. The solutions were filtered and further analyzed to determine total metal concentrations of some selected heavy metals using the AAS, Buck Scientific 210VGP. Standard wavelengths used for Pb, Zn, As, Cd, Cu, Ni, Co and Mn were 283.3nm, 307.6nm, 331.8nm, 326.1nm, 327.4nm, 341.5nm, 352.7nm and 279.5nm respectively. The wavelength accuracy and detection limit of the analyses were $\pm 0.2\text{nm}$ and $0.01 \mu\text{g ml}^{-1}$ respectively according to ISO 26845:2008 and ICH 2000. Other parameters measured both in the field and laboratory include pH, temperature, total dissolved solid (TDS), and electrical conductivity (EC). In addition, petrographic microscope studies on thin sections were conducted on the igneous rock samples according to ASTM C1721, Standard Guide. For petrographic study, instrumentation analysis was done using Hillquist Thin Section Machine, Model 1005 & 1010 and Olympus petrological microscope. While sample preparations were carried out at the Departments of Geology, Civil Engineering, Biochemistry and Industrial Chemistry at the Nnamdi Azikiwe University, Awka, AAS analysis was conducted at the Projects Development Institute (PRODA), Federal Ministry of Science and Technology, Enugu, and petrographic studies were done at the Geology Department, Kogi State University, Aigba, all in Nigeria.

4. Results and Discussion

Geochemical analysis of soil, stream sediments, groundwater, surface water and ores were conducted to determine concentrations of Pb, Zn, Cu, Cd, Ni, Co, Mn, and As. The results are shown in Tables 1-4. The results reveal that the highest total concentration of heavy metals observed within the area were: for Cu: 2.80 mg/l, Zn: 0.82 mg/l, As: 5.20 mg/l, Cd: 0.33 mg/l, Ni: 2.26 mg/l, Co: 0.54 mg/l, and Pb: 3.40 mg/l (Table 1). According to the World Health Organization (WHO 2004, 2011) guidelines for drinking water quality, acceptable selected elements concentrations are: for Cd: 0.003 mg/l, As: 0.01 mg/l, Mn: 0.4 mg/l, Ni: 0.07 mg/l, Pb: 0.01 mg/l, Cu: 2 mg/l, Zn: 5 mg/l, Co: 0.05 mg/l. This study shows that Pb, Cd, As, and Ni concentrations in water sources are above the acceptable level for drinking water quality and hence, are potential pollutants. Most of Co concentration levels in water sources are within the WHO standard for drinking water, except those around the Pb-Zn mining site within the Iheututu area. The results also show that concentration levels of heavy metals and metalloid are generally higher in surface and groundwater samples than samples from other media. Mn shows highest concentration in the soil whereas Cu shows highest concentration in the stream sediment. As shows highest concentration in groundwater, followed by Pb and Ni.

Table 1. Results of water analysis for Ishiagu area

Location	1	2	3	4	5	6	7	8	9	10	11
Temperature ($^{\circ}\text{C}$)	28.7	27.3	31.0	30.5	30.2	28.7	38.0	28.0	29.0	29.3	27.1
pH	6.05	5.00	6.28	6.36	6.25	6.35	7.07	6.63	6.57	6.95	5.77
TDS (calculated)	170	20	40	100	60	50	20	40	390	1270	230
EC ($\mu\text{s/cm}$)	254	30	60	149	90	75	30	60	583	1896	343
Constituents (mg/l)											
Copper (Cu)	0.40	1.20	0.90	<0.10	0.60	1.10	0.30	1.60	0.70	<0.10	0.10
Zinc (Zn)	<0.10	0.24	0.25	0.25	0.07	0.82	0.04	0.24	0.07	<0.10	0.30
Arsenic (As)	<0.10	0.80	3.00	3.00	2.60	4.00	5.20	2.90	0.80	1.70	2.20
Cadmium (Cd)	0.04	0.18	0.10	0.18	0.22	0.05	0.14	0.23	0.01	0.13	0.19
Nickel (Ni)	0.60	0.29	0.48	1.10	1.11	0.26	0.30	0.44	1.70	<0.10	1.15
Cobalt (Co)	0.03	0.01	0.05	0.08	0.54	0.09	0.04	0.01	0.06	0.01	0.06
Lead (Pb)	1.10	1.10	0.80	1.20	3.00	0.10	0.60	1.10	3.40	0.10	2.00

Locations: 1 = Ngwongwo; 2 = Amaeze Plaza Hotel 1; 3 = Amagu Ikwo; 4 = Iheututu Iyiodu; 5 = Iheututu Ikwo; 6 = Iheututu 1; 7 = Iheututu 2; 8 = Amagu Ivo; 9 = Amaokwe Borehole; 10 = Ngwongwo CRI; 11 = Amaeze Plaza Hotel 2.

Table 2. Result of soil analysis for Ishiagu town

Location	9a	9b	6	7	8	12
Depth Horizon (cm)	15	7	10	5	7	8
Constituents (mg/l)						
Copper (Cu)	0.60	0.60	0.40	0.60	0.20	0.10
Zinc (Zn)	<0.10	0.27	0.07	<0.10	0.07	0.16
Manganese (Mn)	2.23	<0.10	5.84	<0.10	<0.10	2.68
Cadmium (Cd)	<0.10	0.06	0.09	0.010	1.40	0.69
Nickel (Ni)	2.26	0.62	0.15	<0.10	1.40	0.69
Cobalt (Co)	<0.10	0.03	0.10	0.08	0.02	0.09
Lead (Pb)	0.30	<0.1	1.10	0.10	1.10	<0.10
Arsenic (As)	<0.10	<0.10	<0.10	<0.10	<0.10	<0.10

Locations: 9a = Amaokwe CRI; 9b = Amaokwe CRI Intrusives; 6 = Iheututu 1; 7 = Iheututu 2; 8 = Amagu Ivo; 12 = Amaonye.

Table 3. Results of stream sediment analysis for Ishiagu town

Location	1a	13	4a
Constituents (mg/l)			
Copper (Cu)	2.80	1.00	<0.10
Zinc (Zn)	0.12	0.41	<0.10
Manganese (Mn)	<0.10	<0.10	<0.10
Cadmium (Cd)	0.13	0.06	0.07
Nickel (Ni)	1.50	0.54	1.34
Cobalt (Co)	0.02	<0.10	0.05
Lead (Pb)	0.80	1.20	0.70
Arsenic (As)	<0.10	<0.10	<0.10

Locations: 1a = Ngwongwo; 13 = Ike River; 4a = Iheututu Iyiodu

Table 4. Results of tailing analysis for Pb-Zn mine tailings Ihetutu area. TS = Tailings sample

Location	Ihetutu Pb-Zn mine		
Sample code number	TS 1	TS2	TS3
Constituents (mg/l)			
Copper (Cu)	0.20	0.10	1.10
Zinc (Zn)	0.08	0.37	0.06
Arsenic (As)	0.09	<0.10	0.20
Cadmium (Cd)	0.33	0.14	0.12
Nickel (Ni)	0.87	0.64	0.08
Cobalt (Co)	0.08	<0.10	0.03
Lead (Pb)	0.50	0.00	1.70

4.1 Geochemical Analysis

A combination of statistical and geostatistical analysis was applied in processing, analyzing, interpreting and discussing the results. This process involved application of QQPlot, semivariogram and kriging interpolation of contaminants in samples from water medium. Water medium was selected due to its wider distribution and sample representation within the study area.

Statistical analysis of distribution normality of contaminants was conducted. As normally distributed data is expected to plot as a straight line (Zhang, 2006, Masi & Mogan, 2005, Bohling, 2005), the QQPlots to assess normality of Cu, Zn, Cd, As, Ni, Co and Pb in water medium, clearly show normal distribution (Figure 3).

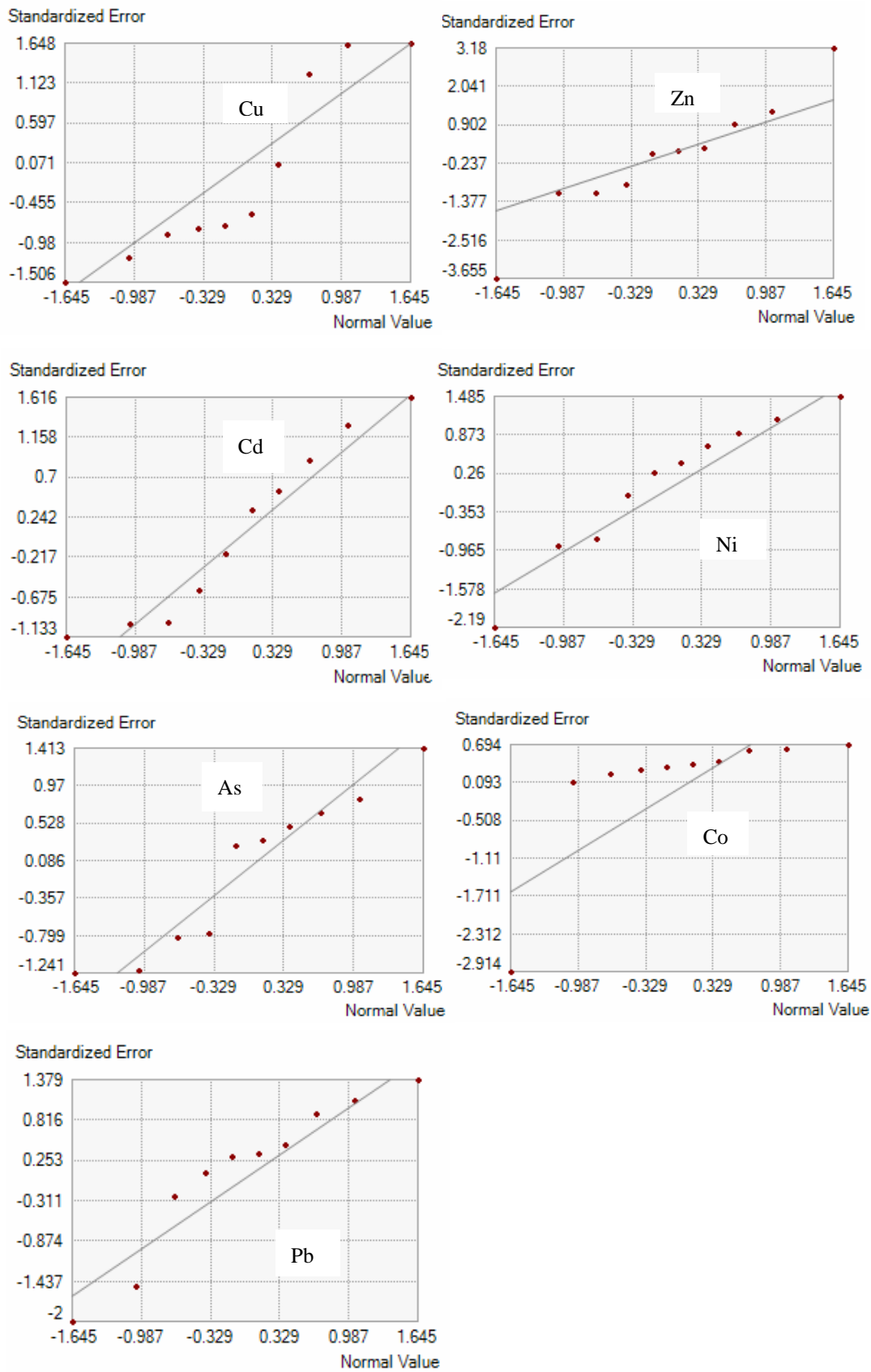


Figure 3. Normal Q-QPlots showing normality of contaminants concentration distribution except Co

From the above plots, Co QQPlot line deviate from 45° and are not very close to being straight, indicating non-lognormal distribution of contaminant. Other elements show normal distribution. Evaluation of spatial semivariance of contaminants concentration pairs with respect to their sample location distances shows mostly Stable Model semivariogram with plume (equidimensional) structure (Figure 4). While majority of the contaminants indicate near zero (0) nugget effect, Pb however, exhibits above one (1) nugget value (Table 5). Generally, there is a spatial correlation of homogeneity at distances larger than the typical sample spacing, without trend of variation, thus, indicating natural ore deposit as primary source of contaminants. The nugget effect can be attributed to a number of factors including the source soil/rock type, environmental mobility effect on elements and impacts of human activities. Taking Pb into consideration, the nugget effect indicates that both natural and anthropogenic factors could have influenced the near even distribution of this contaminant within certain locations of the study area. Anthropogenic activities such as borehole drilling (as observed at Amaokwe) through Pb deposit can trigger migration of Pb from its natural source (ore deposit) into the aquifer system. Similarly, the sills of the contaminants reflect blanket nature, with Cu, Zn and As having sill values of above 1, and show more concentration levels within common location range than Cd, Ni, Co and Pb.

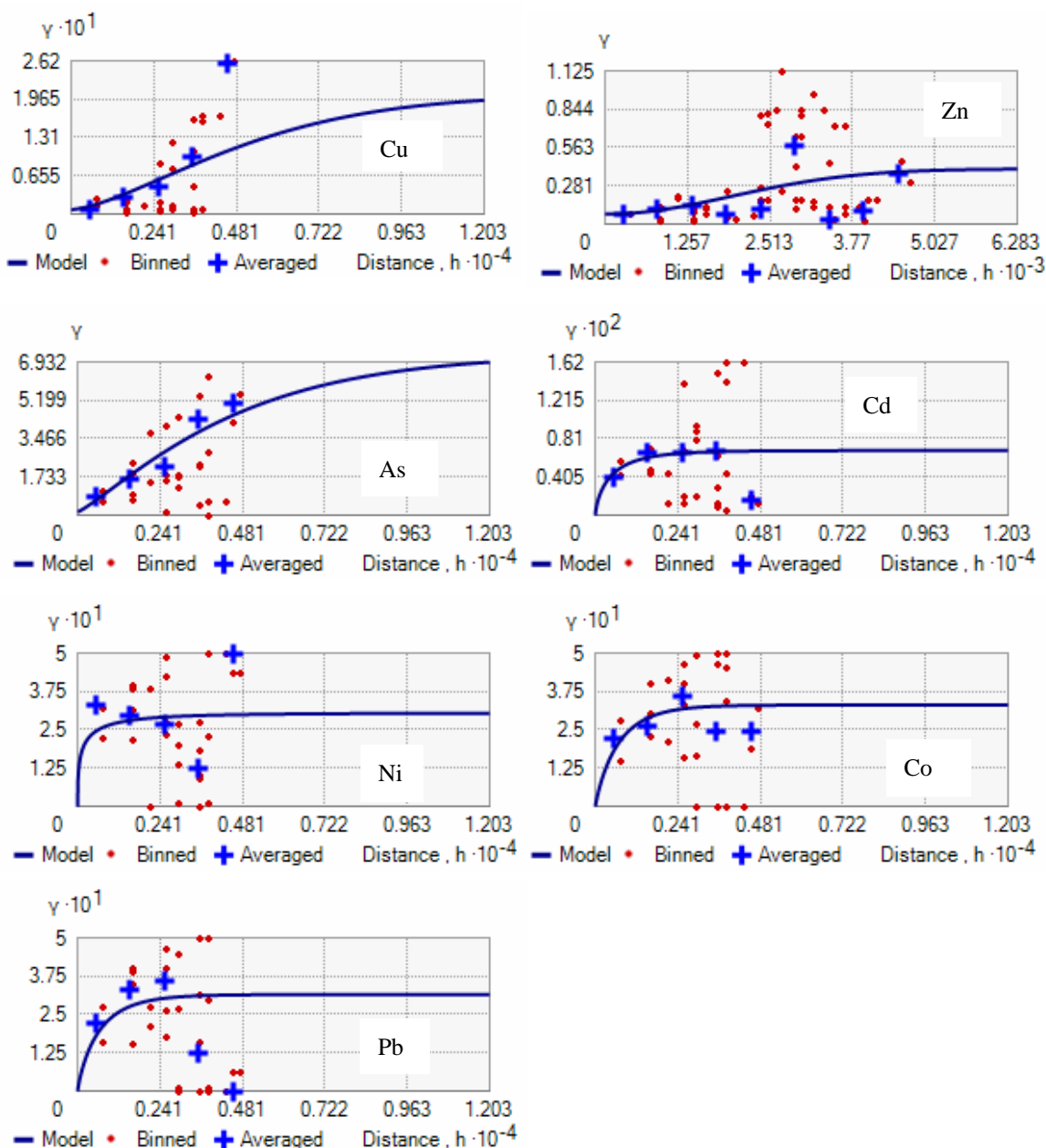


Figure 4. Plots showing semivariogram results of contaminants

Table 5. Semivariogram parameters

Parameter	Model type	Lag size/No of lag	Nugget	Partial sill	Sill	Range
Cu	Stable	523/12	0.07	0.33	2.0	4701
Zn	Stable	1002/12	0.01	0.20	1.5	12034
As	Stable	1002/12	0.16	7.12	1.2	12034
Cd	Stable	1002/12	0.00	0.01	0.7	2312
Ni	Stable	1002/12	0.24	0.00	0.2	12034
Co	Stable	1002/12	0.03	0.00	0.2	12034
Pb	Stable	1002/12	1.23	0.00	0.2	12034

The Ordinary Kriging (OK) method, among the many different kriging techniques, was used in this study due to its simplicity and prediction accuracy when compared with other kriging methods (Issaks & Srivastava, 1989, ESRI, 2010). Ordinary Kriging non-stationary conditions were taken into account by limiting the domain of stationarity to a local neighbourhood and moving it across the study area. After normalization of data for interpolation of some selected heavy metals and metalloid, kriging techniques were used to predict spatial distribution of these metals and metalloid. A variogram was further drawn to select suitable model for fitness on experimental variogram using Less Mean Standardized (MS) value, while best interpolation method was selected using cross-validation and Root Mean Square Standardized (RMSS), as shown in Table 6. GIS (ESRI) ArcMap 10.3.1 was further used to produce geochemical maps (Fig. 5a - h) of heavy metals and metalloid based on this interpolation. The directional influence and trend analysis of geochemical maps from kriging surfaces depict hot-spots of contaminants in the northeast direction. Considering the northwest-southeast flow direction of river distributaries in the area and abundance of igneous rock intrusives associated with hydrothermal deposits (Olubambi et al., 2008, Obiorah & Umeji, 1995) majorly in the northeast area of the site, the folded and faulted anticline features trending northeast with dips of 18° – 36° to the northwest and southeast, the indication is a more natural source than anthropogenic contaminant source.

Table 6. Prediction errors – fitted parameters of the variogram models for heavy metals and metalloid in water samples

Parameter	Prediction error				
	Mean	RMS	ASE	MS	RMSS
Cu	-0.16	0.5	0.39	-0.08	1.13
Zn	-0.03	0.3	0.16	-0.09	1.7
As	-0.01	0.9	1.14	-0.01	0.88
Cd	0.01	0.07	0.08	0.08	0.95
Ni	0.05	0.55	0.52	0.10	1.06
Co	0.01	0.17	0.16	0.08	1.01
Pb	0.11	1.22	1.18	0.09	1.03

Where RMS = root mean square; ASE = average standard error; MS = mean standardized and RMSS = root mean square standardized.

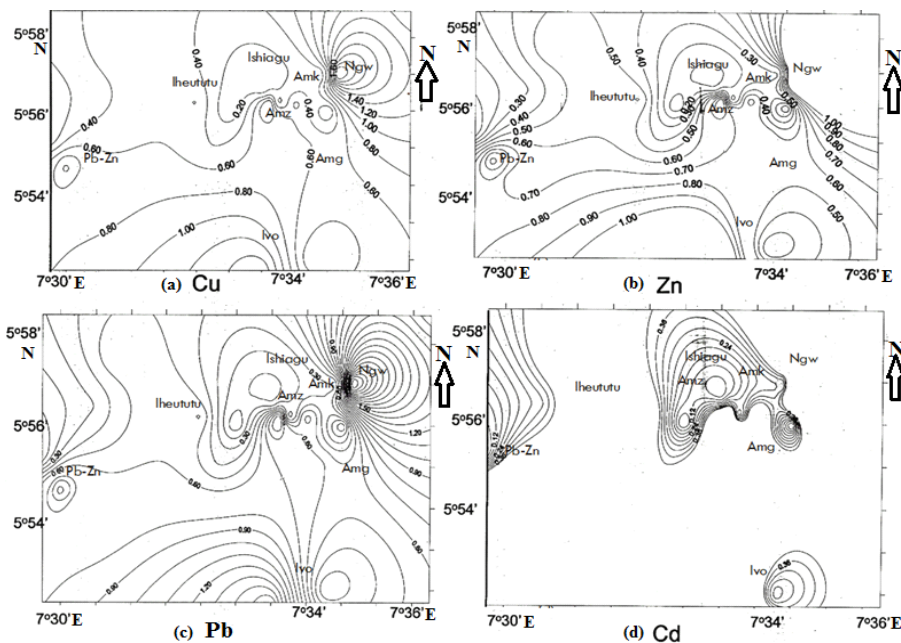


Figure 5 (a-d). Geochemical maps of some selected heavy metals in rock, soil, and water and stream sediment media in Ishiagu town showing dispersion and distribution patterns of the elements. Ngw = Ngwongwo, Amk = Amaokwe, Amg = Amagu, Amz = Amaeze

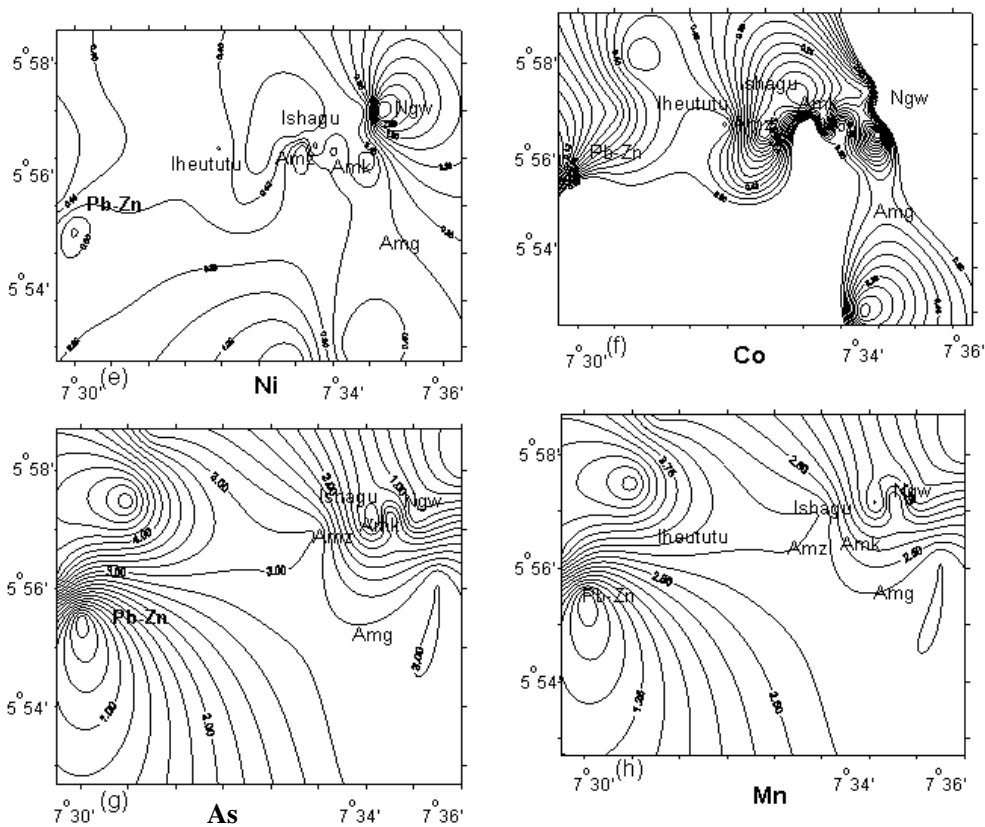


Figure 5 (e-h). Geochemical maps of some selected heavy metals in rock, soil, water and stream sediment media in Ishiagu town showing dispersion and distribution patterns of the elements. Ngw = Ngwongwo, Amk = Amaokwe, Amg = Amagu, Amz = Amaeze

The ASE values show that there is a bit of over estimation of as value as its ASE value is higher than the corresponding RMS value. A cross-validation analysis to examine the validity of fitting models and parameters of semivariogram shows that the predictions do not deviate much from the measured values, as the RMS error and the ASE differences are minimal.

While the MS values are closest to zero, the RMSS values are closest to 1; reflecting a good Kriging mapping surface of the contaminants' homogeneity. Kriging surfaces show significant concentrations of Pb (up to 3.40 mg/l) in both the mineralized (Ikwo Iheututu) and non-mineralized zones (Amaokwe and Amaeze), and indicates that either lead is less mobile than zinc and cadmium in leachates from waste materials and/or that groundwater is seeping through some undiscovered ore deposits in Amaokwe and Amaeze areas. The low concentration of Zn could be attributed to low iron sulfide content, supported by near-neutral zinc – bearing seepage (Iheututu seepage) and massive presence of siderite. These align with Plant et al. (2000), Carla (2002) and Foley (2011, 2002) that Pb and cadmium generally tend to be more mobile in acid soils than in alkaline soils. In addition, near neutralization in leachates occurs from resultant effects of sulphide oxidation and calcite (Wahlstrom et al., 2009, Jiang et al., 2015). The solubility of Pb increases in an acidic environment. In the study area, the pH shows near neutral to alkaline environment as a result of the substantial decrease in the equilibrium carbonate concentration. Thus, the presence of calcium carbonate (Figure 6 & 7) reduces the solubility of Pb, and, in turn, lowers its concentration levels in some areas.

The maps of geochemical distribution of heavy metals as illustrated in Figure 5 show hot spots of heavy metal contamination within Iheututu zone (close to the Pb-Zn mine), with background shale rocks typically containing about 5.20 mg/l As, 1.70 mg/l Pb, 0.37 mg/l Zn and 5.84 mg/l Mn. Cu and Co are evenly distributed in both zones. Cd and Pb increase towards the Iheututu zone, Ni and Zn increase towards the Ishiagu zone while Mn occurs in soil with boulders of the intrusive. In addition, the stream sediments analysis results show that Cu, Zn, Cd, and Pb concentrations increase towards the Ishiagu zone, Ni is evenly distributed in both zones, while Mn is absent. Arsenic concentration is observed to be highest at Amaokwe borehole within the Ishiagu zone, containing about 5.20 mg/l, indicating that As is distributed more within the mineralised zone than the non-mineralised zone.

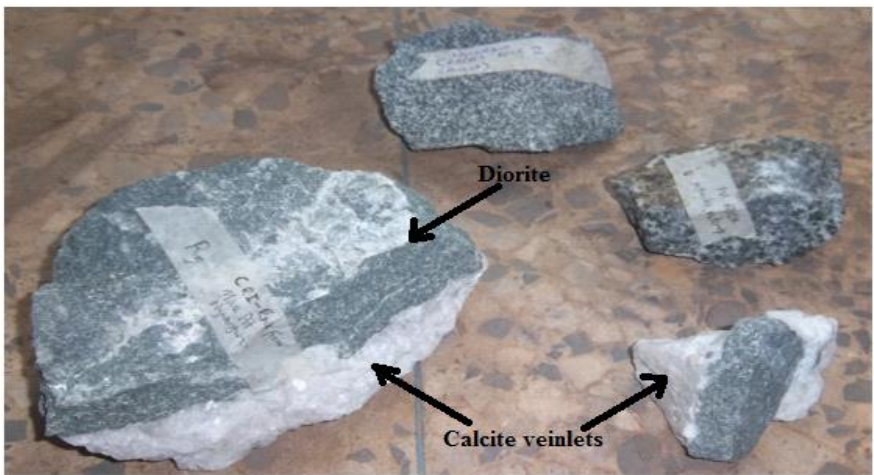


Figure 6. Rock samples from Ishiagu Crushed Rocks Industries. Arrows show calcite veinlets in diorite



Figure 7. Ore (Pb-Zn) tailings with carbonate mineral (calcite) from Pb-Zn mining site at Iheututu area

5. Conclusions

The analysis of the spatial distribution of Pb, As, Cd, Ni and Zn concentrations in surface water, groundwater, stream sediment and soil in the study area showed that those at pollutant levels are Pb, As, Cd, Ni and, to a lesser extent, Cu. Their concentration levels call for remedial measures as groundwater and surface water are the only sources of potable water in the area. Water contained higher concentrations relative to the other media, and the concentrations generally increase downstream. Pb concentration is significantly higher in the groundwater within the Ishiagu (built up) area, even with the absence of Pb-Zn mining activities, than in the other areas. Acidity and Cu concentration are also higher in this area (groundwater pH is up to 5 in places). It is therefore inferred that the Ishiagu area holds particularly higher unexposed and/or undiscovered ore deposits that are the sources of the observed high values. Similarly, the generally high Pb concentration in groundwater relative to surface water in the whole study area suggests possible groundwater flow across relatively deep, unmined and/or undiscovered ore deposits in the study area.

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